

MICROBIAL DEGRADATION OF POLYLACTIC ACID BIOPLASTIC

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Abstract: In modern urban society, plastic has become a major component in many applications. Despite their usefulness, plastics are known for their alarming resistance to biodegradation and therefore pose a huge burden on the environment. The surge in plastic waste has led to the development of biodegradable plastics. Biodegradable plastics are perceived to be one of the sustainable ways to reduce plastic wastes. Among the biodegradable plastics, polylactic acid (PLA), a bio-based plastic, has been widely used for diverse applications. PLA however, is known to be less susceptible to biodegradation and requires longer time to degrade in the natural environment compared to other aliphatic polyesters. Therefore, identification of PLA-degraders is the current focus of research. Potential PLA-degraders have been isolated from various sources, but the number is still limited. Considering temperature as one of the important factors in biodegradation, in this review, we categorize PLA degraders into three groups; thermophilic, mesophilic and psychrophilic, and discuss their relevance to PLA biodegradation. Enzymes associated with PLA degradation are described according to their temperature preference. One of the current limitations is the structural insights into the catalytic mechanism. Discovery of more PLA-degraders and understanding of the mechanism could aid in designing strategies to accelerate the biodegradation process.

Keywords: Biodegradable plastics, sustainability, polylactic acid, biodegradation, PLA-degraders.

Introduction

Accumulation of plastic waste is a current global concern that requires instant and continuous action from various sectors. Although plastics are useful in daily life due to their attractive characteristics such as high durability, tensile strength and flexibility, the growing demand for plastics by various industries makes it difficult to resolve the issue over poor recycling and improper waste management. The food and beverage sector is one of the fast-emerging sectors that is accelerating demand for plastics due to the rising popularity of packaged products. Takeout containers, microwaveable food trays, tableware and soft drink bottles are some of the products heavily consumed in the market. In addition, the unprecedented COVID-19 pandemic has inadvertently contributed to the

increase in plastic wastes due to the rising use of food takeaways and goods deliveries that are plastic packaging-heavy during and even after the movement restrictions imposed by many countries.

Worldwide, plastics in the market are dominated by synthetic-based polymers, mainly petroleum hydrocarbons. Many plastics such as polyethylene (PET), polystyrene (PS), polypropylene (PP), polyethylene terephthalate (PET), and poly vinyl chloride (PVC) have been used in many industrial sectors not only for flexible and rigid packaging but also in building and construction, electrical and electronic products, and housewares. Despite their usefulness, these petroleum-based plastics are known to be resilient to degradation. Their persistence in the environment at the same

state or small fragments (microplastics) for long periods of time poses a serious burden and adverse impact on the environment by harming organisms in various terrestrial and aquatic habitats (Sun *et al.*, 2018; Jiang *et al.*, 2019; Lahive *et al.*, 2019), altering the natural biological and chemical processes and putting the food webs at risk of toxin introduction (Kale *et al.*, 2015). Current treatments for plastic wastes include disposal in landfills, incineration, and chemical treatment and recycling (Gan & Zhang 2019; Moharir & Kumar, 2019). Disposal of plastic waste in landfills and incineration is associated with high energy consumption and generation of secondary pollutants (Webb *et al.*, 2013; Moharir & Kumar, 2019) while recycling is seen to be inefficient and compromising the quality of the polymer. Recent studies are more dedicated to finding solutions to treat plastic waste in a sustainable manner.

One of the promising ways to solve the plastic problem that is plaguing the environment is to consider alternative plastics having these two characteristics; 1) are derived from sustainable materials; 2) can be naturally degraded. On this premise, biodegradable plastics and bioplastics are introduced to the market. Besides Europe, bioplastics have captured the plastic market in Asia-Pacific countries, like China, India, South Korea and Japan, where the demand is steadily increasing, with 80% derived from the packaging industry (Source: Mordor Intelligence). In Malaysia, campaigns and policies to discourage single-use plastics have been implemented, and development of biodegradable plastics has started to grow. Malaysia aims to achieve zero single-use plastics for a cleaner and healthier environment by 2030. Current production of bioplastics in Malaysia is driven by the abundance of agricultural by-products such as palm oil waste and non-food crops that serve as renewable feedstocks.

Taken into account the current state of biodegradable plastics and mode of decomposition by biological treatment, this review focuses on the role of microbes specifically bacteria and their enzymes

in degrading one of the most popular and promising biodegradable plastic, poly (L-lactic acid) (PLA). We highlight the diversity of PLA-degraders in terms of temperature preference and adaptation (thermophilic, mesophilic and psychrophilic) to provide useful information on the relevance of each bacterial category for isolation and identification of more PLA degraders and PLA depolymerases. At the end of the review section, we include some current understanding on PLA hydrolysis mechanism to shed light on the knowledge at molecular level and increase understanding of enzymatic degradation.

Biodegradable Plastics

Biodegradable plastics are compostable in the environment. Commonly used biodegradable plastics are poly (L-lactic acid) (PLA), poly (butylene succinate-co-adipate) (PBSA), poly (butylene succinate) (PBS) and polycaprolactone (PCL). PBS and PCL are grouped as biodegradable because they can be degraded by microorganisms although they are petroleum-based plastics. PHB and PLA, on the other hand, are synthesized from biological-based resources. These sustainable, eco-efficient products made up of renewable agricultural and biomass feedstock can attract markets and potentially outperform the petroleum-based products that are currently prevalent (Mohee *et al.*, 2008). Bioplastics have attracted considerable interest due to their properties which are comparable to synthetic polymers and can potentially replace conventional synthetic plastics. Although the term 'biodegradable' has become appealing to the mass market and consumers, yet, in the natural environment, complete degradation will take years or decades, and potentially destructive small particles are generated in the process (Kubowicz & Booth, 2017). Biodegradation involves a combined action of abiotic and biotic factors (Lucas *et al.*, 2008; Sivan *et al.*, 2011). Fragmentation of polymers into smaller pieces is suggested to be the starting process of degradation and the plastic is considered biodegradable only if it is

consumed and metabolized by microorganisms in the subsequent process (Mohee *et al.*, 2008). The action of exoenzymes secreted by microorganisms is important to degrade the larger polymers before it can be incorporated into the microbial cells (Danso *et al.*, 2019). In this review, we focus on PLA-type bioplastic because of its current wide usage in diverse applications.

Poly(lactic) Acid Bioplastic

Among the bioplastics, PLA presents the most promising biomass plastic with good mechanical properties and processability (Jem & Tan, 2020). PLA has the prerequisites of sustainability by being both biodegradable and bioplastic. The attractiveness of PLA in terms of its biocompatibility, high transparency, stiffness and thermoresistance (Elsawy *et al.*, 2017) makes it useful for many purposes. PLA has been widely utilized as packaging films, containers, bottles with short shelf-life and single-used products (Jem & Tan, 2020). Lactic acid, the monomer of PLA, is made from the fermentation of renewable sources, such as starch, corn and sugar. Lactic acid is a chiral molecule, existing as L- and D-lactic acid isomers. The production of PLA from renewable sources is energy saving. PLA can also be chemically synthesized by ring-opening polymerization of lactide or direct-condensation polymerization of lactic acid. Despite its interesting characteristics and great potential as a substitute for petroleum-based plastics, PLA has some limitations, including its biodegradability. The biodegradation of PLA has gained considerable attention because of its widespread use and disposal into the environment. Between the two stereoisomers of PLA; poly(L-lactic) PLLA and poly(D-lactic) PDLA, biodegradation of PLLA is more frequently reported compared to PDLA. In general, under typical composting facilities, PLA is degraded at high temperature (>50 °C) (Kawai *et al.*, 2011) and requires months to be fully degraded. Due to this drawback, there is an urgent need for an efficient and eco-friendly method to degrade PLA.

Microbial Degradation of Biodegradable Plastics

The use of microorganisms is a promising solution to the plastic disposal issue and has been a focus of recent research. Intriguingly, the interactions between microorganisms, plastics and environmental factors may influence the environmental fate of terrestrial and marine plastics. In general, the entire decomposition of biodegradable plastics through bacterial and fungal activity occurs at slow rates (Muhonja *et al.*, 2018). The biodegradation rate is influenced by many factors including the molecular weight of the polymer (Tokiwa & Calabia, 2004; Urbanek *et al.*, 2020), the shape of the specimen, either in powder or film, (Yang *et al.*, 2005) or particle size and surface area (Chinaglia *et al.*, 2018). High molecular weight polymers are more recalcitrant to degradation and are very resistant to attack by enzymes (Danso *et al.*, 2019). Natural degradation of polymers usually starts with the abiotic influence, such as photodegradation and thermo-oxidative degradation followed by the action of microbes (Webb *et al.*, 2013). These abiotic processes affect the chemical structure of the polymer chains. Apart from the properties of the polymer, degradation of biodegradable plastics is also determined by additives added to the final consumer products and the environment that they terminate. Nevertheless, because this biological method is a promising, potentially efficient, safer and more environmentally friendly approach than chemical and physical methods, such as landfill disposal and incineration, it is, therefore, important to discover more microorganisms that can degrade plastics, which can further be manipulated to increase the efficiency of biodegradation.

Microorganisms have developed many unique features that allow them to adapt to unfavourable conditions and evolve in contaminated environments. Microorganisms isolated from specific or extreme environments have shown many potentials for biotechnology applications and are often exploited for bioremediation and biodegradation processes.

Based on the literature, plastic-degrading microbes have been extensively isolated from soil and compost compared to marine or aquatic environments. Also, it is worth noting that specific species may have the capability to degrade specific types of plastic or several types of polyesters. For example, *Paenibacillus amylolyticus* isolated from soil could degrade various types of polyesters including PLA, PBS, PBSA, PCL and PES and showed preference to degrade low-molecular weight polymers (Teeraphatpornchai *et al.*, 2003). Yagi *et al.* (2014) identified unique eubacteria participating in anaerobic biodegradation of PCL and PLA at 37°C suggesting that some microbes are substrate specific.

In plastic-contaminated areas, biodegradable plastics are a source of carbon and energy (Chinaglia *et al.*, 2018; Decorasi *et al.*, 2019) likely when microbial nutrients are scant and limited. Under aerobic conditions, microorganisms assimilate the carbon chain of the brittle, small pieces plastics and mineralize it to carbon dioxide or use it to form the biomolecules (Zheng *et al.*, 2005; Chinaglia *et al.*, 2018). In general, degradation of polyesters by microorganisms has been suggested to involve mainly three sequential steps: 1) surface colonization or biofilm growth to physically break down the polyesters, 2) action of microbial exoenzymes to produce fragments of polyesters, and 3) assimilation of intermediates/molecules from the small fragments into bacterial metabolism (Mueller, 2006; Lucas *et al.*, 2008).

Degradation of Biodegradable Plastics by Extremophiles

Extremophiles are one of the emerging sources of plastic-degrading microorganisms. There is a growing interest in extremophiles because of their ability to produce a wide array of useful exoenzymes (Di Donato *et al.*, 2019). Extremophiles are characterized by their physical and chemical parameters that allow them to grow optimally. On the basis of temperature, extremophiles are divided into two main

categories: thermophiles and psychrophiles. Thermophiles are microorganisms that can grow at temperatures above 60°C, typically 60-80°C, while hyperthermophiles are those that survive at temperatures above 80°C. At elevated temperatures, thermophiles have high metabolism and produce enzymes that are highly stable (Haki & Rakshit, 2003). It is well documented that thermophiles demonstrate substantial potential of degrading various kinds of environmental organic pollutants largely because at high temperature solubility and bioavailability of organic pollutants are enhanced (Margesin & Schinner, 2001). Thermophiles have been widely reported in biodegradation of hydrocarbons such as petroleum hydrocarbons and currently perceived to be valuable in biodegradation of plastics, another carbon-based pollutant.

An increasing number of studies have found many potent extremophiles from various sources that have the ability to degrade plastics (Hadad *et al.*, 2005; Michaud *et al.*, 2007; Hanphakphoom *et al.*, 2014, Shah *et al.*, 2015; Dang *et al.*, 2018). For example, Shah *et al.* (2005) isolated a bacterium identified as *Ralstonia* sp. MRT-TL from a hot spring that is capable of degrading 50% of PCL film within 10 days. The weight loss of the film is likely due to the breakdown of the long PCL chain into low molecular weight compounds. Çolak *et al.* (2005) reported a novel hot spring thermophile, *Anoxybacillus gonensis* G2A, that can decompose PHB film at 60°C. The weight loss of P(3HB) films was 6.9%, 13.4% and 24.0% observed at 24, 48 and 72 h, respectively. A novel thermophilic bacterial strain known as *Bacillus* sp. BCBT21 is capable of decomposing three different types of plastic bags – HL, VHL, and VN1 – with different chemical nature after 30 days (Dang *et al.*, 2018).

Meanwhile, 70% of earth is covered by sea and serves as a reservoir of psychrophilic microorganisms at temperatures below 15°C. Psychrotrophs (cold-tolerant) have a broader growth temperature range and are capable of growing optimally at temperatures

above 15°C (Morita, 1975; Moyer & Morita, 2007). Psychrophiles and psychrotrophs play an important role in the biodegradation of contaminants in cold environments, such as oceans, sea and polar regions. In contrast to high-temperature regions, contaminants may persist longer in cold environments because of the extreme condition and low bioavailability. In a marine environment, plastic waste accumulation, especially microplastics, has negatively impacted the ecosystem. But the information on biodegradation of plastics by aquatic microbes remains limited. A study conducted by Dussud *et al.* (2018) revealed that degradation of biodegradable polymers (artificially aged OXO and poly (3-hydroxybutyrate-co-3-hydroxyvalerate, PHBV) in an aquatic setting (aquariums with natural circulating seawater) was 30 times higher than that of non-degradable polymers. Members of *Neptuniibacter* sp., *Phaeobacter* sp., and *Roseobacter* sp. were dominantly found in the early colonization, growth and maturation phases, respectively. However, the role of each species needs to be further described.

Some of the microorganisms are also categorized as polyextremophiles, where they demonstrate the ability to survive at several extreme conditions. Piezophiles, microorganisms that can thrive in cold and high hydrostatic pressure in deep-sea environments, have also been reported capable of degrading aliphatic polyesters. In a study by Sekiguchi *et al.* (2010), four strains of piezophilic and piezotolerant bacteria (strains CT01, CT12, JT01 and JT04) isolated from deep-sea environments demonstrated high potential to degrade biodegradable PCL-type plastic. While many marine extremophiles are capable of producing enzymes with unique features (Di Donato *et al.*, 2019) and could be of great interest to

biotechnological applications, those reported to be involved in plastic degradation are relatively few and many are still unexplored. Therefore, it is apparent that much has to be done to identify potential plastic degrading bacteria in order to overcome the plastic waste problem and sustain both life on land and below water.

PLA-degrading Bacteria

Among all aliphatic polyesters, PLA is often associated with slow degradability due to low abundance of microbes that are able to decompose PLA in the environment (Tokiwa & Calabria, 2004). Also, PLA decomposition is challenging because of its resistance to microbial attack under ambient settings (Farah *et al.*, 2016). This poses an alarming threat as many countries are shifting to bioplastic usage. The slow degradation could lead to unnoticed accumulation of plastic waste, be it large or small fragments, before the degradation is complete.

Bacteria are known to be the key players in many biodegradation processes. Current research on PLA-degrading bacteria suggests that most PLA-degrading bacteria include members of Actinobacteria and predominantly belong to the *Amycolatopsis* genus (Butbunchu & Pathom Aree, 2019). Because of their high occurrence as PLA-degraders, natural biodegradation of PLA is suggested to be influenced by the role of the members of this genus. In addition, members from the families of *Bacillaceae*, *Pseudomonadaceae*, *Paenibacillaceae*, *Flavobacteriaceae*, *Xanthomonadaceae* have also been reported (Table 1). Because PLA biodegradation is influenced by temperature, in the following section, we divide the PLA-degraders into three categories; thermophiles, mesophiles, and psychrophiles and discuss their relevance to PLA biodegradation.

Table 1: Several thermophilic, mesophilic and psychrophilic bacteria capable of degrading PLA

Category	Genus/Species	Source	Optimal Growth/ Cultivation Temperature (°C)	Type of PLA/ Form of PLA	Reference
Thermophilic	<i>Bacillus smithii</i>	Garbage fermentor	40-65 (optimal at 55)	PLLA	Sakai <i>et al.</i> (2001)
	<i>Bacillus stearothermophilus</i>	Soil	60	PDLA	Tomita <i>et al.</i> (2003)
	<i>Actinomadura</i> sp. strain T16-1	Thai forest soil	50	PLA film	Sukkhum <i>et al.</i> (2009)
	<i>Laceyella sacchari</i> strain LP175	Forest soil	50-55	PLLA	Hanphakphoom <i>et al.</i> (2014)
	<i>Bacillus licheniformis</i>	Compost	50	PLA	Prema & Palempalli (2015)
Mesophilic	<i>Amycolatopsis</i> sp. strain K104-1	Soil	37	High MW PLA	Nakamura <i>et al.</i> (2001)
	<i>Paenibacillus amylolyticus</i> strain TB-13	Soil	30	PLA emulsion	Teeraphatpornchai <i>et al.</i> (2003)
	<i>Amycolatopsis orientalis</i> ssp. <i>orientalis</i>	Culture collection	30	PLA film	Li <i>et al.</i> (2008)
	<i>Amycolatopsis thailandensis</i> sp. nov.	Soil	25-37	PLA emulsion	Chomchoei <i>et al.</i> (2011)
	<i>Pseudomonas</i> sp. strain DS04-T	Activated sludge	37	PLLA	Wang <i>et al.</i> (2011)
	<i>Aneurinibacillus migulanus</i>	Soil	30-40.5	PLA film	Chaisu <i>et al.</i> (2012)
	<i>Bacillus amyloliquefaciens</i> MS2	Compost	NR	PLA film	Prema & Palempalli (2014)
	<i>Pseudomonas</i> sp. MYK1 <i>Bacillus</i> sp. MYK2	Digester sludge	30	PLA	Kim <i>et al.</i> (2017)
	<i>Chryseobacterium</i> sp. <i>Sphingobacterium</i> sp. <i>Pseudomonas aeruginosa</i> S2 <i>Pseudomonas aeruginosa</i> S3	Compost	30	PLA film	Satti <i>et al.</i> (2017)

	<i>Stenotrophomonas pavanii</i> CH1 <i>Pseudomonas geniculata</i> WS3	Soil from agricultural areas, sanitary landfill sites wastewater sludges	20-45	PLA	Bubpachat <i>et al.</i> (2018)
	<i>Bacillus pumilus</i> (B12)	Agricultural soil	21-37	High molecular weight PLA film	Bonifer <i>et al.</i> (2019)
Psychrophilic	<i>Pseudomonas</i> sp. and <i>Rhodococcus</i> sp.	Arctic soil	28	PLA	Urbanek <i>et al.</i> (2017)

Note: NR - Not Reported

Thermophilic PLA-degrading Bacteria

Thermophilic bacteria have been widely reported capable of degrading PLA compared to their mesophilic and psychrophilic counterparts. The focus of isolating thermophilic bacteria is based on the notion that natural degradation of PLA (PLLA) in landfills and recycling process by enzymatic method takes place at high temperature (Hanphakphoom *et al.*, 2013). Besides, effective biodegradation of bioplastics was observed between 70% and 90% at a temperature range of 58–65°C (Ruggero *et al.*, 2019). Most thermophilic PLA-degrading bacteria have been isolated from various sources including agricultural soil, garbage fermentor, forest soil, activated sludge and compost (Sakai *et al.*, 2001; Prema & Palempalli, 2015). A thermophilic PLA-degrading bacterium was isolated from a garbage fermentor which operated at 55-60°C by Sakai and co-workers (2001). This bacterial strain grew at 45-60°C and exhibited optimal growth temperature at 55°C. However, under aerobic conditions at its optimal temperature, the growth was poor. A soil-isolate closely related to species *Geobacillus thermocatenulatus* was able to grow at 60°C on medium supplemented with 0.5% PLA film (Tomita *et al.*, 2004). Sukkhum *et al.* (2009) used a clear zone method to isolate PLA-degrading actinomycetes from various forest soils at high temperature. Among the 13 isolates,

strain T16-1 showed the largest clearing zone (3.6 cm) and highest PLA degradation activity (22 U/mL) in the liquid culture. The strain was identified as *Actinomadura keratinilytica* strain T16-1, a member of *Thermomonosporaceae* family. Meanwhile, Hanphakphoom *et al.* (2013) isolated eleven strains of thermophilic poly(L-lactide)-degrading bacteria from forest soils that showed positive clearing zone on emulsified PLLA agar plate. Strain LP175 which was identified as *Laceyella sacchari* demonstrated the highest PLLA degradability at 50°C for 4 days. A thermophilic *Bacillus licheniformis* isolated from compost of a plastic rich environment showed a significant degradation of PLA at 50°C (Prema & Palempalli, 2015).

Mesophilic PLA-degrading Bacteria

While thermophilic PLA-degrading bacteria are advantageous especially for biodegradation in composting facilities that usually operate at 40-70°C, application of PLA-degraders that work at moderate and low temperatures can be economical in terms of energy consumption. Furthermore, biodegradation of PLA in nature, given the conditions in most settings, is mainly driven by the activity of mesophilic bacteria (Pattanasuttichonlakul *et al.*, 2018). In this sense, mesophilic and psychrophilic bacteria capable of breaking down polyesters can provide more

options to strategize efficient ways to rapidly decompose bioplastics in different environments. Soils from landfill, compost and agricultural lands, as well as wastewater sludges are sources of mesophilic PLA-degrading bacteria. A higher probability to obtain PLA-degrading bacteria is observed in soils exposed or contaminated with plastic wastes, such as sanitary landfill and wastewater sludges compared to agricultural soils (Bubpachat *et al.*, 2018). A soil *Actinomyces* identified as *Amycolatopsis thailandensis* sp. nov. (strain CMU-PLA07) was reported to be able to degrade PLA. The bacteria had optimal growth temperature at 25-37°C and no growth was observed at 45°C (Chamchoei *et al.*, 2011). A-PLA degrading mesophile was isolated from compost and based on 16S rDNA sequence analysis, the species was identified as *Bacillus amyloliquefaciens* MS2 (Prema & Palempalli, 2014). Satti *et al.* (2017) tested the capability of four isolates from compost to degrade PLA at 30°C by supplementing the culture with PLA as the sole carbon source. All four strains (*Chryseobacterium* sp. strain S1, *Sphingobacterium* sp. strain S2, and *Pseudomonas aeruginosa* strain S3 & S4) demonstrated PLA degradation ability at ambient temperature. Bubpachat and co-workers (2018) reported two PLA-degrading bacteria isolated from soil and wastewater sludge in an attempt to search for new efficient mesophilic PLA-degrading strains. The two strains, *Stenotrophomonas pavanii* and *Pseudomonas geniculate* WS3 were able to grow at temperatures of 25-40°C and degrade PLA at 30°C. Interestingly, *P. geniculate* WS3 demonstrated efficient degradation activity toward beverage cup PLA and PLA by more than 90% of weight loss in the non-sterilized soil mixture under thermophilic conditions for 12 days (Pattanasuttichonlakul *et al.*, 2018). Decorasi *et al.* (2019) isolated four *Actinomyces* strains (designated as SO1.1, SO1.2, SNC, and SST) from soils of different geographical areas. All four strains were able to grow at temperatures below 45°C (28 and 37°C) and thus are grouped as mesophiles. Strain SNC isolated from soil close to a city road showed the best degradative

capabilities at 30°C by inducing 36% weight loss of PLA film in soil extract medium with 0.1% gelatin.

Psychrophilic PLA-degrading Bacteria

Since plastic wastes can be found in numerous parts of the Earth, including cold environments, it is likely that more polyesters-degrading microorganisms can be discovered. Psychrophilic microorganisms can be found in low temperature environments, including the deep sea, mountains and polar regions (D'Amico *et al.*, 2006). Other sources include high-altitude mountains, glaciers or natural caves, while oceans provide the largest psychrophilic reservoir (Mergesen & Feller, 2010). Although current studies on psychrophiles capable of degrading plastic are scarce, the potentials shown by this group of microorganisms should not be neglected. Psychrophiles may be employed for biodegradation in marine ecosystems, during winter in temperate regions or other cold ecosystems. Similar to its higher counterparts, the attractiveness of psychrophiles is due to their unique enzymes. Since a large amount of plastic debris has been entering the oceans for many decades and changing the indigenous ecosystem, marine bacteria have evolved the capability to colonize and interact with plastic surfaces. This interaction, which is influenced by many factors, such as substrate type, surrounding environment, geographical location and seasonal variation of environmental parameters causes the formation of biofilm (Kirstein *et al.*, 2019).

Biodegradable plastics, such as PCL, PBSA and P3HB are proven biodegraded in the marine environment and the rate depends on the variety of organisms in the sea, and physical and chemical effects (Nakayama *et al.*, 2019). In the same study, PLLA was found to be biodegraded at a very slow rate, even after 4 weeks of treatment using seawater. A study reported psychrophilic microorganisms displaying microbial activity towards PLA. Urbanek and co-workers (2017) isolated psychrophilic bacteria from Arctic soil in an attempt to identify Arctic microorganisms with best biodegradability features. Out of 113 bacterial isolates, 45.13% of the bacteria grew

on 0.1% emulsified PLA at 28°C after 2-3 days during the initial screening, however PLA was the hardest to degrade compared to other tested polymers, PCL and PBSA. The resilience of PLA against degradation has been suggested to be related to the low humidity when cultured on PLA-emulsified plates (Urbanek *et al.*, 2017).

PLA Depolymerases

The bio-based polymer is preferentially attacked by microorganisms. Since plastic is made up of polymer with high molecular weight, this polymer is unable to cross the cell membrane (Pattanasuttichonlakul *et al.*, 2018). During the depolymerization process, microorganisms secrete extracellular enzymes that are responsible for breaking down complex polymeric structures to short monomers (Palmisano & Pettigrew, 1992). Recent studies are focusing more on characterizing the PLA-degrading enzymes or known as PLA depolymerase. Hydrolases such as protease, lipase, esterase and cutinase are identified to play a chief role in the biodegradation of PLA. The fact that PLA is an aliphatic polyester means that the mechanism of degradation of PLA is primarily by the action of hydrolases on the ester bond (Li *et al.*, 1994; Panyachanakul *et al.*, 2019). According to Lenz (1993), there are two modes of hydrolytic cleavage: 1) exo-attack, where the hydrolysis takes place at the polymer chain terminus and results in small oligomers or monomers that can be incorporated into the cells, or 2) endo-attack, where hydrolysis occurs somewhere along the polymer chain and the products cannot be assimilated by the cells.

Among the PLA-degrading hydrolases, serine proteases or protease-like enzymes have been earlier reported to be associated with PLA degradation, where proteinase K was the first PLA-degrading enzyme reported (Williams, 1981). Esterase and lipase belong to the same family and both capable of degrading PLA but are distinguished by their substrate preference. Esterase acts on shorter chain fatty acid esters, thus targeting the low molecular weight PLA, whereas lipase preferentially hydrolyses high

molecular weight PLA (Urbanek *et al.*, 2020). Cutinase is an enzyme of mainly fungal origin and acts on cutin substrate. Cutinase and lipase share similar endo-type action in biodegradation of polyester by cleaving the ester bond between alcohol and carboxylic acid (Shi *et al.*, 2020). Unlike lipase, cutinase hydrolyses esters with shorter chain length ($C < 10$) and lacks interfacial activation, such as that observed in lipase family.

To date, most reported biodegradation of PLA involves poly (L-lactic acid) PLLA (Sukkhum *et al.*, 2009; Chomchoei *et al.*, 2011; Hanphakphoom *et al.*, 2014) whereas reports on hydrolysis poly (D-lactic acid) PDLA by PLA depolymerase are very limited. In a study conducted by Kawai *et al.* (2011), on the hydrolytic activities of protease, lipase and PLA-degrading enzymes to different types of PLA; (poly (L-lactic acid) (PLLA) and poly (D-lactic acid)), they concluded that protease activity is more specific to PLLA while lipase prefers PDLA as substrate.

Thermostable PLA Depolymerases

In many bio-based industrial processes, extremophilic enzymes offer many attractive improvements to reaction efficiency. Of these, thermophilic enzymes have long been considered as a valuable biocatalyst for high-temperature reactions primarily due to their robust stability and increased resistance to proteolysis at elevated temperatures. Enzymes are considered thermostable when they can function above 55°C; usually in the range of 60-80°C and can originate either from thermophilic or mesophilic microorganisms. At temperatures below 40°C, the activity will typically diminish (Vieille & Zeikus, 2001). The application of thermostable enzymes is seen not limited only to industrial processes, but they are also useful in bioremediation. Evidently, there has been a considerable interest in identification of novel thermostable PLA-degrading enzymes. Table 2 summarizes several thermostable bacterial PLA-depolymerases that have been reported. A purified extracellular fibrinolytic serine protease is able to degrade PLA at temperatures

of 55-60°C. This 24-kDa protein hydrolyses high molecular weight PLA and the maximum degradation was observed at pH 9.5 (Nakamura *et al.*, 2001). A crude wild-type esterase from *Bacillus smithii* works optimally at 60°C on PLA and other fatty acid esters (Sakai *et al.*, 2001). Akutsu-Shigeno *et al.* (2003) purified and characterized a recombinant PLA depolymerase which functions at 45-55°C. This alkaline depolymerase is capable of decomposing a variety of biodegradable polyesters including PLA. PLA with lower molecular weight showed high degradability. Besides PLA, this enzyme could hydrolyse triolein and tributyrin and various p-nitrophenyl alkyl esters, especially butyl ester. Three thermostable PLAases (PLAase I, II, and III) was reported by Li *et al.* (2008). The enzymes originating from *Amycolatopsis orientalis* ssp. *orientalis* hydrolysed PLA and casein, but the activity was inhibited by aprotinin. A thermostable serine protease was identified from *Actinomadura* sp. strain T16-1. Hydrolysis of PLA was optimum at 70°C (Sukkhum *et al.*, 2009).

Meanwhile, Prema & Palempalli (2015) described a PLA depolymerase from *Bacillus licheniformis* that is active at 50-60°C. Apart from PLA, the hydrolytic activity was observed in the presence of gelatin and casein as substrates. A PLA depolymerase from *Pseudomonas tamsuii* TKU015 exhibited optimum activity at 60°C and retained its activity by 50% at 80°C. It is worth mentioning that the bacterial strain was identified as a mesophilic *Pseudomonas*. Hydrolytic activity determination suggested that lipase and protease activity were not present. This enzyme could represent a novel PLA depolymerase that acts on the ester bond between lactate units but not on some common lipase and protease substrates (Liang *et al.*, 2016).

Mesostable and Cold Active/Adapted PLA Depolymerases

In contrast to thermostable enzymes, mesophilic (mesostable) and psychrophilic (cold active/adapted) enzymes usually work best at

temperature ranges of 25 to 50°C and 5 to 25°C, respectively (Vieille & Zeikus, 2001). Several mesophilic and psychrophilic PLA depolymerases have been reported (Table 3). A PLA depolymerase which identified as elastase-like protease was reported from *Amycolatopsis* sp. strain K104-1. The purified enzyme obtained from extracellular fraction consists of 238 amino and the mature enzyme encodes for a protein with a size of 20,904 Da. Although the optimal temperature was not specifically mentioned, degradation test using PLA emulsion was conducted at 37°C. A decrease in emulsion turbidity was recorded within 3 to 6 hours and complete disappearance of PLA spot (measured by thin layer chromatography) was observed after 6 hours (Matsuda *et al.*, 2005). Bulpachat and co-workers (2018) isolated two mesophilic bacteria producing protease and PLA-degrading enzymes that hydrolyse PLA at 30°C. *Stenotrophomonas pavanii* CH1 produced a protease that has optimal pH of 7.0-7.5. *Pseudomonas geniculata* WS3 produced the highest levels of PLA-degrading enzyme at pH 8.0. Recently, an esterase from *Pseudomonas aeruginosa* strain S3 capable of depolymerizing PLA films was purified with a molecular mass of approximately 34 kDa. This enzyme was stable at wide temperature and pH ranges, with highest stability at 30°C and pH 7.0 (Noor *et al.*, 2020).

Structural flexibility of enzymes from psychrophilic organisms allows them to exhibit 10 times higher activity at low and moderate temperatures compared to their mesophilic homologues (Morgesin & Feller, 2010; Cavicchioli *et al.*, 2011), thus, enzymes from psychrophiles can efficiently break down PLA at cold temperatures. To date, there is little information on cold-active/adapted PLA-depolymerases. One of the few psychrophilic PLA-degrading enzymes reported is ABO2449 from *Alcanivorax borkumensis*. This enzyme is suggested to be a cold-adapted esterase by its ability to retain 32% of its maximal activity at 5°C and hydrolyses soluble naphthyl- and nitrophenyl-esters of C2-C4. Furthermore, the cold-adapted property was confirmed by its low thermostability ($T_{agg} = 32.3 \pm 0.5^\circ\text{C}$)

Table 2: Several reported thermostable bacterial PLA-depolymerases/PLA-degrading enzymes

Enzyme	Fraction	Molecular Weight (kDa)	Optimum Temperature (°C)	Optimum pH	Substrate/ Inhibitor	Reference
Serine protease	Purified wild-type (culture supernatant)	24.0	55-60	9.5	High MW PLA	Nakamura <i>et al.</i> (2001)
Serine protease	Purified wild-type (culture supernatant)	30.0	70	10.0	PLA, Gelatin, Suc-(Ala)3-pNA Inhibited by EDTA, PMSF and diisopropyl fluorophosphates	Sukkhum <i>et al.</i> (2009)
PLLA degrading enzyme (esterase)	Purified wild-type (culture supernatant)	62.5	60	5	PLLA and fatty acids esters	Sakai <i>et al.</i> (2001)
PLA depolymerase	Purified recombinant	22.0	45-55	10	PLA, PBS, PBSA, PCL, PES, Tributyrin, Triolein	Akutsu-Shigeno <i>et al.</i> (2003)
PLAase I	Purified wild-type	24.0	60	9.5	PLA, casein	Li <i>et al.</i> (2008)
PLAase II	(culture supernatant)	19.5	50	10.5	Inhibited by aprotinin and PMSF (PLAase activity)	
PLAase III	(culture supernatant)	18.0	60	9.5		
PLA depolymerase	Purified wild-type (culture supernatant)	34.0	50	8.5	PLA, triolein, poly(b-hydroxybutyrate), and poly(ε-caprolactone)	Wang <i>et al.</i> (2011)
PLLA degrading enzyme (serine protease)	Purified wild-type (culture supernatant)	28.0	60	9	PLLA, casein and gelatin	Hanphak-phoom <i>et al.</i> (2014)
PLA depolymerase	Purified wild-type (culture supernatant)	44.0	50-60	6-7	PLA, casein, gelatin	Prema & Palempalli (2015)
PLA depolymerase	Purified wild-type (culture supernatant)	58.0	60	10	PLA, Fibrinogen, tributyrin	Liang <i>et al.</i> (2016)
PLA degrading enzyme	Crude wild-type	30.0	60	8	PLA	Panya-chanakul <i>et al.</i> (2019)

Note: NR - Not Reported

Table 3: Several reported mesostable and cold-adapted bacterial PLA depolymerases/PLA-degrading enzymes (including PLA depolymerases tested at moderate temperature)

Enzyme	Fraction	Molecular Weight (kDa)	Optimum/ Tested Temperature (°C)	Optimum pH	Substrate	Reference
Esterase	Purified recombinant	29.8	37°C for PLA hydrolysis	NR	PLA and PET	Ribitsch <i>et al.</i> (2012)
Esterase ABO2449	Purified recombinant	NR	30-37 (retain 32% of its maximal activity at 5 °C)	pH 9.5-10	PLA, Monoester substrates, tributyrin (C4)	Hajighasemi <i>et al.</i> (2016)
Esterase	Purified wild-type	34	30	pH 7	PLA film	Noor <i>et al.</i> (2020)
PLA depolymerase (serine protease)	Purified wild-type (culture supernatant)	72.5	NR (Tested at 37 °C)	NR	PLA, casein, gelatin	Prema & Palempalli (2014)
PLA depolymerase (elastase-like protease)	Purified recombinant (culture supernatant)	24.2 (mature 20.9)	37	NR	PLA	Matsuda <i>et al.</i> (2005)
Protease & PLA degrading enzyme	Crude wild-type	NR	NR (Tested at 30 °C)	pH 7.5-pH 8.0	PLA & gelatin	Bubpachat <i>et al.</i> (2018)

Note: NR - Not Reported

and the ability of the bacteria to grow at 4°C (Hajighasemi *et al.*, 2016).

Although PLA degrading enzymes acting at moderate and low temperature have been reported, the efficiency of biodegradation is accompanied by some limitations. Studies have shown that biodegradation of PLA in mesophilic conditions is slower compared to thermophilic conditions. The change in the molecular weight of the polymer occurs at a slow rate in both aerobic and anaerobic settings, likely due to the lack of capability to hydrolyse the high molecular weight polyesters, such as PLA and PCL (Yagi *et al.*, 2014; Bubpachat *et al.*, 2018). According to a study by Ribitsch *et al.* (2012) the polymer chain must be flexible to provide higher accessibility for enzymatic attack and

hydrolysis, thus, reactions at high temperatures are favourable. Nevertheless, PLA degradation in mesophilic conditions is deemed to be important as it represents the natural degradation in most environments.

Current Understanding on PLA Depolymerase Structure and Mechanism

In general, the literature suggests that PLA depolymerase falls in the group of hydrolases. Therefore, structural investigations are important to unveil the mechanism of PLA hydrolysis by PLA depolymerases. Identification of the active site and key residues involved in the catalysis may assist in the design and manipulation of the enzyme to overcome the limitations of enzyme-mediated PLA degradation. Several

studies have examined and analysed the amino acid sequences and functional roles of specific residues in bacterial PLA depolymerases.

Lipase, esterase and cutinase share the canonical α/β hydrolase fold with Ser-His-Asp as the catalytic triad. Meanwhile, protease, specifically serine protease, has been reported to be involved in the hydrolysis of polyesters (Lim *et al.*, 2005). The presence of serine, which acts as a nucleophile, is critical to drive the hydrolytic reaction. Hydrolytic mechanism by serine protease occurs in two-sequential steps, acylation and deacylation (Hedstrom, 2002). In general, the reaction begins when the hydroxyl group of a serine residue attacks the substrate scissile bond forming the first tetrahedral intermediate, acyl enzyme and releasing the first product. A second intermediate is formed when a water molecule attacks the acyl enzyme and the second product is released.

The gene encoding PLA depolymerase from *Paenibacillus amylolyticus* strain TB-13 was cloned and over-expressed in *Escherichia coli*. The recombinant lipase-type PLA depolymerase displayed a conserved pentapeptide Ala-His-Ser-Met-Gly such as that present in *Bacillus* lipase (Akutsu-Shigeno *et al.*, 2003). Similarly, the presence of serine residue in the catalytic site of PLA depolymerase from *Bacillus amyloliquefaciens* MS2 was confirmed by a complete inhibition when treated with serine inhibitors, phenylmethylsulfonyl fluoride and aprotinin (Prema & Palempalli, 2015).

Matsuda *et al.* (2005) suggested that the depolymerization of PLA by an elastase-like PLA depolymerase follows the catalytic mechanism of the chymotrypsin family. Based on sequence analysis, residues H74, D111, and S197 are proposed to be the catalytic triad of the elastase-like PLA depolymerase. The oxygen of nucleophilic Ser195 (chymotrypsin numbering) attacks the peptide bond or ester bond and generates an acyl intermediate and liberates the first product. Subsequently during the deacylation step, covalent intermediate is hydrolysed and active enzymes are regenerated when the second product is released.

An extensive screening of 90 purified microbial α/β hydrolases by Hajjighasemi *et al.* (2016) led to the identification of two uncharacterized proteins ABO2449 from *Alcanivorax borkumensis* and RPA1511 from *Rhodospseudomonas palustris* that are able to completely hydrolyse solid PLA. The crystal structure of RPA1511 revealed that the catalytic triad consists of Ser114, Asp242 and His270. The catalytic Ser114 and several hydrophobic residues, including Ile245 and Val249, forms the substrate acyl-binding pocket of RPA1511. The mechanism was proposed based on classic Ser hydrolase catalysis comprising acylation and deacylation steps. Several residues (Gln172, Leu212, Met215, Trp218, and Leu220) were also identified to be the key in PLA hydrolysis by RPA1511.

Future Perspective and Challenges

PLA is a promising sustainable material that potentially serves as a substitute for petroleum-based plastics. The emergence of PLA in the market raises the issues related to its slow degradability. The current challenge is to improve the biodegradation rate in order to avoid problems related to plastic pollution from worsening. Microorganisms are known as the key players in polymer biodegradation, however the diversity of PLA-degrading bacteria and known PLA depolymerases are still rather limited. Although the number of reports is increasing, the importance and limitation of different groups of bacteria has not been extensively discussed. The review summarizes the current status and gaps in PLA biodegradation, especially in exploration of PLA degraders and their relevant enzymes. Given different environments and conditions, thermophilic, mesophilic and psychrophilic bacteria are important for biodegradation at various settings. Also, it is evidenced that several thermostable PLA-degrading enzymes are produced by mesophilic bacteria. In addition, further studies need to be directed to understanding the biodegradation mechanism at molecular level to allow engineering of PLA-depolymerases with robust activity and broad

substrate specificity. Other aspects that could be described include the complex interactions between enzymatic degradation and factors that could accelerate the biodegradation of PLA in both laboratory setup and natural environment.

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